

# KINETICS OF ANAEROBIC PROCESSES

M. Kryłów\* and B. Tal- Figiel\*\*

\* Institute of Water Supply and Environmental Protection, Cracow University of Technology,  
ul. Warszawska 24, Kraków, Poland  
(E-mail: [malkryl@vistula.wis.pk.edu.pl](mailto:malkryl@vistula.wis.pk.edu.pl))

\*\* Department of Engineering and Technology of Chemistry, Cracow University of Technology,  
ul. Warszawska 24 , Kraków, Poland

## ABSTRACT

In the paper, description of anaerobic treatment of municipal wastewater, according to the proposed kinetic model, has been presented. The experimental results of microbial growth predictions showed the greatest resemblance to the models of Monod, Teisser, Contois as well as Chen and Hashimoto.

## KEYWORDS

Anaerobic biodegradation, kinetics model, methanogenesis, acetogenesis, acidogenesis.

## INTRODUCTION

During the research, the author focused on anaerobic treatment of municipal wastewater. So far, this subject as well as a complete description of microbial growth kinetics, have not been thoroughly investigated.

To fully understand and describe the kinetics of anaerobic biodegradation process the experiments were conducted in continuous flow reactor:

Three phases of the anaerobic wastewater treatment process were investigated:

- metanogenesis,
- acetogenesis
- acidogenesis.

## EXPERIMENTAL PROCEDURE

The experimental installation consisted of a glass reactor placed in a thermostat. The reactor diameter was 0.08 m, and its height 0.023 m. The reactor volume was 0.001 m<sup>3</sup>. The reactor content was mixed with a magnetic stirrer.

The reactor was fed with anaerobic activated sludge extracted from the UASB reactor. The biomass concentration in the reactor was 26 g dry solids/l and was comparable with the concentration in the UASB reactor.

The anaerobic reactor was seeded with anaerobic activated sludge taken out from the full-scale UASB reactors, operating at the wastewater treatment plants. The sludge in the laboratory scale reactor had a floc structure.

To assure the stability and consistency of the process parameters, throughout the operating cycle, the synthetic wastewater were used in the experiments.

During the course of experiments, the following parameters were changed:

- actual retention time in the reactor,
- influent wastewater concentration,
- temperature 25 °C, 30°C, 35 °C.

In each series of experiments only one parameter (the actual retention time in the reactor) varied. The other parameters such as concentration of COD in the influent to the reactor and temperature remained constant.

During the experiments the following process parameters were measured (according to the Polish Standards):

- pH of the influent and effluent (determined with the accuracy down to 0.01 (PN-72/C-04540/01)
- temperature inside the reactor,
- COD in the influent and effluent to the reactor (PN-74/C-04578.03 oraz 5220B wg Standard Methods 17<sup>th</sup> 1989),
- biomass concentration at different reactor levels and in the effluent (metodą wagową wg PN-72/C-04559.02),
- biogas content (gas chromatography),
- content of carboxylic acids (VFA) in the influent and effluent (gas chromatography) (Kryłów i Lach 1998)].

In addition, the basic rheological parameters of the anaerobic activated sludge were determined using reowiskozymeter HAAKE RS 75.

### PROCESS KINETIC MODEL

The model can accommodate the successive phases of the anaerobic reaction and describe the following mechanism:



...



where : X - biomass,  
 S - substrate,  
 P - product

Based on the above relationships the microbial growth for the entire process can be described as:

$$\mu_C = \frac{dC_X}{C_X \cdot dt}, \quad (4)$$

where:  $C_X = C_{X1} + C_{X2} + \dots + C_{Xn}$ . (5)

It was assumed that the fraction of the particular types of microorganisms in the biomass remains constant. Therefore, a set of kinetic equations can be written as:

$$-\frac{dC_S}{dt} = \mu_1(C_S) \cdot C_{X_1} = r_1 \quad (6)$$

$$\frac{dC_{P_1}}{dt} = Y_{P_1X} \cdot \mu_1(C_S) \cdot C_{X_1} = Y_{P_1X} \cdot r_1 \quad (7)$$

$$\dots$$

$$\frac{dC_{P_n}}{dt} = Y_{P_nX} \cdot \mu_n(C_{P_{n-1}}) \cdot C_{X_n} \quad (8)$$

$$r_1 = r_2 = K = r_n = f(C_S) \quad (9)$$

A mass balance for the individual intermediate product in this system can be expressed as:

$$\frac{dC_{P,n}}{dt} = Y_{P_n,X} \cdot \mu_n(C_{P,n-1}) \cdot C_{X_n} - \frac{\mu_{n+1}(C_{P,n})}{Y_{X,P_n}} \cdot C_{X_{n+1}} \quad (10)$$

An increase of a concentration of those particular microorganisms, which participate in each phases of the process is very difficult to measure and therefore the following relationships have been assumed:

$$C_{X_1} = \varphi_1 \cdot C_X \quad (11)$$

$$C_{X_2} = \varphi_2 \cdot C_X \quad (12)$$

$$\dots$$

$$C_{X_n} = \varphi_n \cdot C_X, \quad (13)$$

where:  $\varphi_1, \varphi_2, \dots, \varphi_n$  – mass fraction of biomass elements.

Substituting equations (11) – (13) in equation (4) and rearranging the latter, the following relationship can be obtained:

$$\mu_C = \mu_1 \cdot \varphi_1 + \mu_2 \cdot \varphi_2 + K + \mu_n \cdot \varphi_n. \quad (14)$$

The relationship describes the overall microbial growth rate of the process as the sum of the bacterial growths observed during each individual stage of the process.

$$\mu_C = \sum_{i=1}^n \mu_i \cdot \varphi_i, \quad (15)$$

where:  $\mu_i$  – specific microbial growth rate in a single individual stage of the process.

If the microorganism growth in each individual stage follows the Monod's kinetics then the overall specific microbial growth rate can be expressed as:

$$\mu_C = \sum_{i=1}^n \left( \varphi_i \cdot \left( \mu_{\max_i} \cdot \frac{C_{S_i}}{K_{S_i} + C_{S_i}} \right) \right). \quad (16)$$

To verify the mechanism assumed for anaerobic biodegradation of wastewater, the mechanism of VFA conversion to methane, with acetic acid as an intermediate product, was investigated. Based on the assumed mechanism, the following set of equations was obtained, each one describing the rate of the particular reaction:

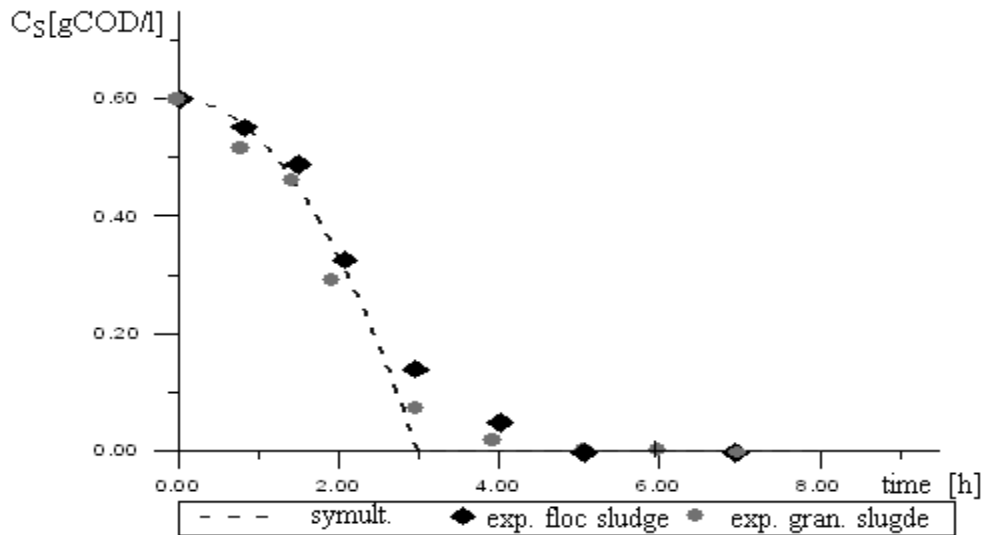
$$\frac{dC_X}{dt} = Y_{XS} \cdot k_1 \cdot C_S \cdot C_X + Y_{XP_p} \cdot k_2 \cdot C_{P_p} \cdot C_X \quad (17)$$

$$\frac{dC_S}{dt} = -k_1 \cdot C_S \cdot C_X \quad (18)$$

$$\frac{dC_{P_p}}{dt} = -k_2 \cdot C_X \cdot C_{P_p} \quad (19)$$

$$\frac{dC_{P_k}}{dt} = Y_{P_k P_p} \cdot k_2 \cdot C_{P_p} \cdot C_X + Y_{P_k S} \cdot k_1 \cdot C_S \cdot C_X \quad (20)$$

As the initial values  $C_{S0}$ ,  $C_{P_p0}$ ,  $C_{P_k0}$ ,  $C_{X0}$ , the values obtained during the experiments were assumed. The experimental values fitted a simulation curve with an error lower than 15%. Hence, it can be stated that the proposed mechanism of substrate biodegradation and intermediate products generation is correct (Fig. 1,2, 3, 4).



**Figure.1.** Substrate concentration versus time

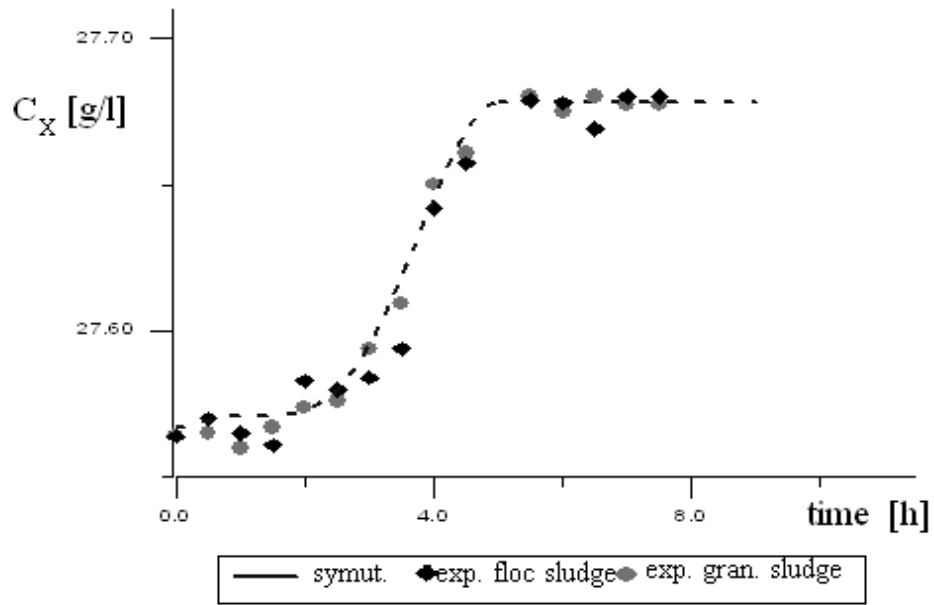


Figure 2. Biomass concentration versus time

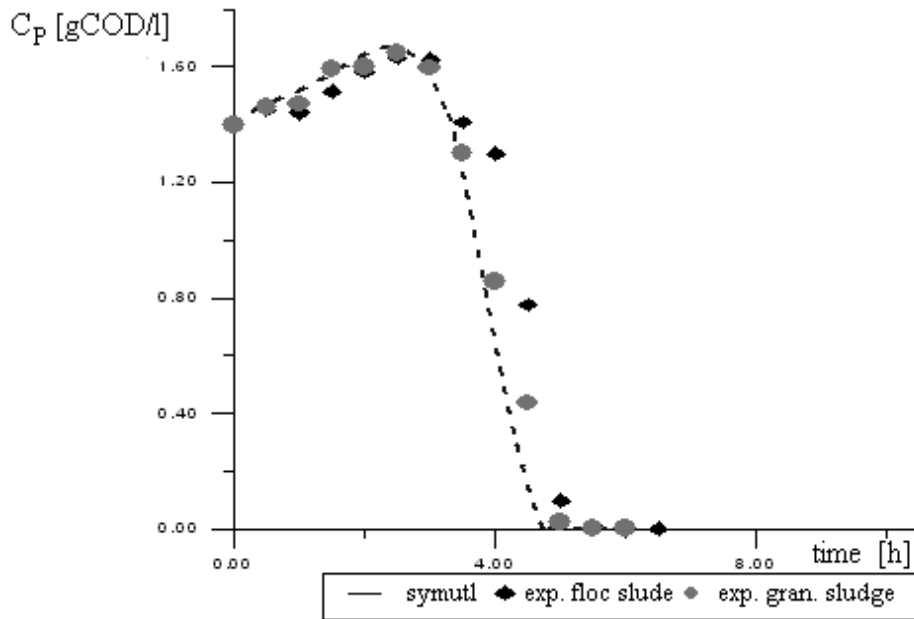
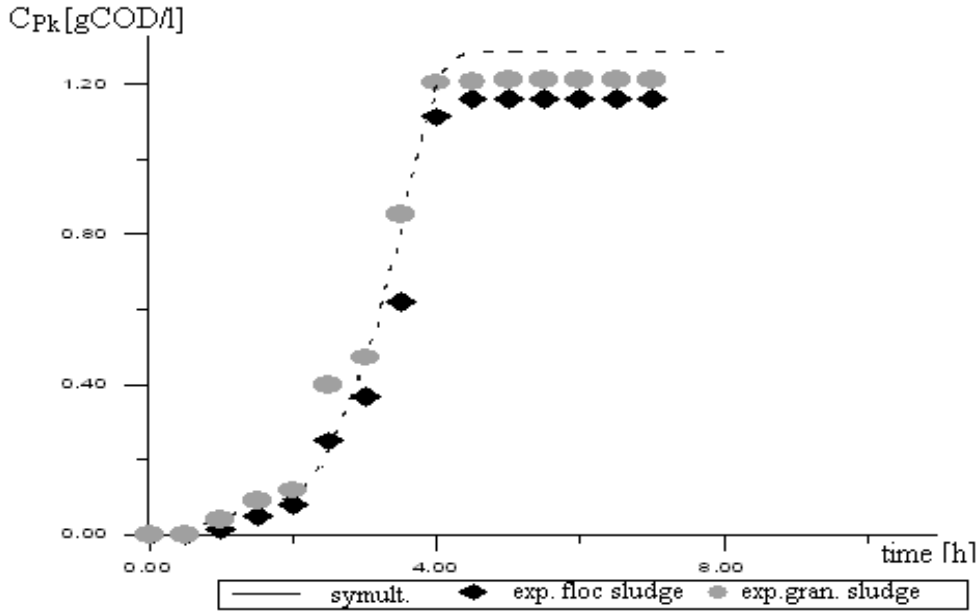


Figure 3. Intermediate product concentration versus time

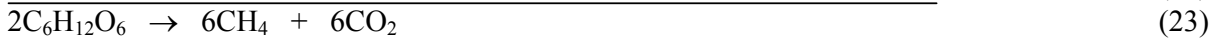
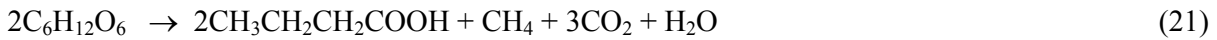


**Figure 4.** Final product concentration versus time

To estimate the values of the kinetic parameters the experimental results were compared with different kinetic models. The best fit was obtained for the Monod, Teisser Contois as well as Chen and Hashimoto models (error less than 10%).

The following example of biodegradation of glucose can serve as an illustration of application of the proposed model of microbial growth in anaerobic wastewater treatment.

Typical conversion of glucose can be described with the set of the following reactions:



The results from the experimental series, with a mixture of propionic and butyric acid as a substrate, were compared with the theoretical models.

Breakdown of glucose to methane gas proceeds as a two-stage process. In the first intermediate stage, propionic and butyric acids are produced, while in the second intermediate stage acetic acid is produced. In this case, the general Monod's equation can be expressed as:

$$\mu = \left[ \varphi_1 \cdot \left( \mu_{\max 1} \cdot \frac{C_{S1}}{K_{S1} + C_{S1}} - b_1 \right) + \varphi_2 \cdot \left( \mu_{\max 2} \frac{C_{S2}}{K_{S2} + C_{S2}} - b_2 \right) + \varphi_3 \cdot \left( \mu_{\max 3} \frac{C_{S3}}{K_{S3} + C_{S3}} - b_3 \right) \right] \quad (24)$$

The paper presents a kinetic model of anaerobic wastewater biodegradation. Following the Monod's kinetics, the kinetic parameters were determined both in graphical and analytical way. The parameters were determined for different temperatures (25°C, 30°C, 35°C) and for different stages of the anaerobic digestion process. The results are presented in Tables 1.

**Table 1.** Kinetic parameters of the Monod's equation for low strength wastewater.

| Substrate                 | Process        | Temp<br>·<br>°C | Sludge   | $Y_{XS}$<br>[g SS. /<br>gCOD] | $K_S$<br>[gCOD<br>/dm <sup>3</sup> ] | $\mu_{max}$<br>[1/h] | $b$<br>[1/h] | $q_P$<br>[m <sup>3</sup> CH <sub>4</sub> /<br>gCOD <sub>rem</sub> ] | $q_{smax}$<br>[gCOD/<br>g SS·h] |
|---------------------------|----------------|-----------------|----------|-------------------------------|--------------------------------------|----------------------|--------------|---|---------------------------------|
| Acetic acid               | Methanogenesis | 25              | floc     | 0.044                         | 0.0175                               | 0.0003               | 0.00001      | 0.15  | 0.146                           |
| VFA (without acetic acid) | Acetogenesis   | 25              | floc     | 0.46                          | 0.0192                               | 0.00043              | 0.00001      | 0.16  | 0.026                           |
| Hydrocarbons (glucose)    | Acidogenesis   | 25              | floc     | 0.05                          | 0.25                                 | 0.073                | 0.001        | 0.165   | 0.194                           |
| Acetic acid               | Methanogenesis | 25              | granules | 0.044                         | 0.02                                 | 0.0008               | 0.00001      | 0.17  | 0.16                            |
| VFA (without acetic acid) | Acetogenesis   | 25              | granules | 0.046                         | 0.034                                | 0.0096               | 0.00001      | 0.182   | 0.0287                          |
| Hydrocarbons (glucose)    | Acidogenesis   | 25              | granules | 0.05                          | 0.125                                | 0.08                 | 0.001        | 0.188   | 0.212                           |
| Acetic acid               | Methanogenesis | 30              | floc     | 0.045                         | 0.01                                 | 0.00038              | 0.0001       | 0.165   | 0.012                           |
| VFA (without acetic acid) | Acetogenesis   | 30              | floc     | 0.047                         | 0.02                                 | 0.00057              | 0.0001       | 0.17  | 0.015                           |
| Hydrocarbons (glucose)    | Acidogenesis   | 30              | floc     | 0.044                         | 0.25                                 | 0.0741               | 0.001        | 0.175   | 0.03                            |
| Acetic acid               | Methanogenesis | 30              | granules | 0.045                         | 0.01                                 | 0.00038              | 0.0001       | 0.165   | 0.012                           |
| VFA (without acetic acid) | Acetogenesis   | 30              | granules | 0.047                         | 0.02                                 | 0.00057              | 0.0001       | 0.17  | 0.015                           |
| Hydrocarbons (glucose)    | Acidogenesis   | 30              | granules | 0.044                         | 0.25                                 | 0.0741               | 0.001        | 0.175   | 0.03                            |
| Acetic acid               | Methanogenesis | 35              | floc     | 0.054                         | 0.035                                | 0.00045              | 0.00001      | 0.172   | 0.0317                          |
| VFA (without acetic acid) | Acetogenesis   | 35              | floc     | 0.047                         | 0.023                                | 0.00063              | 0.00001      | 0.179   | 0.039                           |
| Hydrocarbons (glucose)    | Acidogenesis   | 35              | floc     | 0.055                         | 0.25                                 | 0.0004               | 0.001        | 0.185   | 0.338                           |
| Acetic acid               | Methanogenesis | 35              | granules | 0.054                         | 0.31                                 | 0.00085              | 0.00001      | 0.176   | 0.07                            |
| VFA (without acetic acid) | Acetogenesis   | 35              | granules | 0.047                         | 0.0375                               | 0.0001               | 0.00001      | 0.183   | 0.039                           |
| Hydrocarbons (glucose)    | Acidogenesis   | 35              | granules | 0.055                         | 0.35                                 | 0.084                | 0.001        | 0.196   | 0.352                           |

## SUMMARY

Based on the experimental results of the biodegradation kinetics the following conclusion can be formulated:

1. The Monod's kinetics is convenient to describe the individual stages of the process.
2. The structure of anaerobic activated sludge had no impact on the type of kinetic description of glucose breakdown (acidogenesis) to lower carboxylic acids (propionic and butyric acids) as well as their breakdown to acetic acid (acetogenesis). These anaerobic digestion reactions can be most accurately described by the kinetic formulas of Monod, Contois as well as Chen and Hashimoto.

In the process design and modeling, kinetics of different stages of anaerobic biodegradation of low strength wastewater, in reactors with sludge flocs, can be illustrated with the equations of Monod or Chen and Hashimoto.

## LITERATURE:

- APHA, AWWA & WPRC, (1989) *Standard Methods for the Examination of Water and Wastewater*, 17<sup>th</sup> Edition, American Public Health Association Washington DC
- Arcand Y., Chavarie C., Guiot R. (1949) *Dynamic Modelling of the Population Distribution Anaerobic Granular Biofilm* Water Sci. Technol. 30, (12), 63 – 73
- Birjukow W.W., Kanterie W.M. (1985) *Оптимизация периодических процессов микробиологических синтез*, Moskwa, Nauka.
- Buraczewski G., „*Fermentacja metanowa*”, PWN, Warszawa, (1989)
- Chen Y.R., Hashimoto A. G. (1980) *Substrate Utilization Kinetic Model for Biological Treatment Prozesse*, Biotechnol. Bioengn. 22, 2081-2095 (1980)
- Kryłów M., Lach A. (1998) *Metody analityczne do oznaczania lotnych kwasów tłuszczowych (LKT)*, Materiały na seminarium szkoleniowe pt *Metody oznaczania wskaźników zanieczyszczeń organicznych w wodzie i ściekach*, Kraków,
- Kryłów M. Tal-Figiel B. (1999) *Biodegradacja skostężonych ścieków w reaktorze UASB* Zeszyty Naukowe Politechniki Łódzkiej s. Inż. Chem. i Proc 226, 55-62
- Kryłów M. (2000) *Biodegradacja niskostężonych ścieków w reaktorze UASB*” Rozprawa Doktorska Politechnika Krakowska
- Malina J. F. Jr, Pohland F. G. (1992) *Design of Anaerobic Processes for the Treatment of Industrial and Municipal Wastes*”, Technomic Publishing Inc., Pensywalnia U.S.A., Water Quality Management Library –7
- PN-72/C-4559/02 *Woda i ścieki Badanie zawartości zawiesin Oznaczenie zawiesin ogólnych, mineralnych i lotnych metodą wagową*”
- PN-74/C-4578/03 *Woda i ścieki Oznaczenie chemicznego zapotrzebowania tlenu (ChZT) metodą dwuchromianowa*
- PN-90/C-04540/01 *Woda i ścieki Badanie pH, kwasowości i zasadowości. Oznaczenie pH wód i ścieków o przewodności elektrolitycznej właściwej 10  $\mu$ S/cm i powyżej metodą elektrometryczną*”